

You win some, you lose some

- compensating the loss of green space in cities taking heterogeneous population characteristics into consideration



You win some, you lose some – compensating the loss of green space in cities taking heterogeneous population characteristics into consideration

Jonas Nordström^{a,b,c} and Cecilia Hammarlund^a

^a Agrifood Economics Centre, Lund University, Sweden

^b Department of Food and Resource Economics, University of Copenhagen, Denmark

^c Department of Business, Economics and Law, Dalarna University, Sweden

Abstract

The increased urbanization and human population growth of recent decades have resulted in the loss of urban green spaces. One policy used to prevent the loss of urban green space is ecological compensation. Ecological compensation is the final step in the mitigation hierarchy; compensation measures should thus be a last resort, after all opportunities to implement the earlier steps of the hierarchy have been exhausted. The compensation should balance the ecological damage, aiming for a “no net loss” of biodiversity and ecosystem services. In this study, we develop a simple model that can be used as tool to study the welfare effects of applying ecological compensation when green space is at risk of being exploited, both at an aggregate level for society and for different groups of individuals. Our focus is on urban green space and the value of the ecosystem service – recreation – that urban green space provides. In a case study we show how the model can be used in the planning process to evaluate the welfare effects of compensation measures at various sites within the city. The results from the case study indicate that factors such as population density and proximity to green space have a large impact on aggregate welfare from green space and on net welfare when different compensation sites are compared against each other.

Key words: Urban green space, Ecological compensation, Recreational value, Wellbeing, Utility, Welfare effects, Distributional effects

JEL classification: D60, Q26, Q50, Q57, R52, R58

1. Introduction

The increased urbanization and human population growth of recent decades have resulted in loss of urban green spaces (Wu et al. 2019 and McKinnely 2002). Today, more than half of the global population lives in urban areas, a proportion that is expected to increase to about two thirds by 2050 (United Nations). Urban green space supports biodiversity and provides critical ecosystem services (Tzoulas et al., 2007 and Colding, 2011) such as recreation, relaxation, aesthetic values, and regulation of the microclimate. Additional factors have also contributed to the loss of urban green space, e.g., conflicting government policies, real estate development (Wu et al 2019) and lack of financial support (Colding et al. 2020).

The concept of ecological compensation can be used in the planning and development process for urban areas to prevent the loss of urban green space and environmental degradation. Ecological compensation is the fourth and final step of the so-called mitigation hierarchy (ten Kate et al., 2004; BBOP, 2012). According to this hierarchy, the developer should primarily strive to prevent and minimize losses of biodiversity and ecosystem services by first planning to (1) avoid, (2) minimize and (3) restore biodiversity or remedy the negative environmental impact on site before (4) compensation can become relevant. Compensation measures should thus be applied as a last resort, after all opportunities to implement the earlier steps of the hierarchy have been exhausted. When compensation is implemented, the measures should balance the ecological damage, aiming for a “no net loss” with full compensation for all ecological damage (e.g. Gibbons and Lindenmayer, 2007; Bull et al., 2016). Ecological compensation has increasingly been adapted, and today it is applied in many countries around the world (Moilanen and Kotiahe, 2018, Bull et al., 2015, Bull and Strange, 2018).

Another key term in ecological compensation is additionality. Additionality entails that the gains at the offset area, the area where the compensation is done, must represent an additional and quantifiable improvement over the offset area’s current and future baseline condition. Without generating these additional gains, losses to ecosystem services and biodiversity will not be offset, resulting in a net loss of social welfare.

Although no net loss is a strict requirement for ecological compensation, it is possible to compensate for environmental degradation at different spatial scales, both in close proximity to the impact area or more distant from the impact area, at other sites within the city, region or nation (Moilanen and Kotiahe 2018, Bull et al. 2015). This flexibility implies that the

compensation may be achieved at a lower cost, and that the target no net loss can be reached in a cost-effective way. Sometimes it may for example be difficult to find offset areas in close proximity to the impact area. It may also be more difficult to reach the offset area, e.g. due to barrier effects from infrastructure investments in highways/railroads for which larger compensation is needed to achieve no net loss. For example, to obtain the same amount of recreational values and well-being for residents living in close vicinity to an impacted area, the additional green space at the offset area may have to be larger than the green space that has been lost.

Spatial re-location of green space will also result in distributional effects, however. Some individuals will lose out from the re-location, whilst others will benefit from it.

As urban green spaces do not tend to be evenly distributed among household- and socio-economic groups (de la Barrera et al. 2016; Wolch et al. 2014), ecological compensation can contribute to a more even distribution of green space among socio-economic groups, but also to a more uneven distribution. Whether ecological compensation will contribute to a more even/uneven distribution of green space among socio-economic groups depends on the socio-economic characteristics of the population surrounding the impact and offset area. To evaluate compensation at various offset areas and the distributional effects, a transparent tool that can be used to carry out the analysis is of great value (BBOP, 2009). In this study, we present a simple tool that can be used to carry out such an analysis. In the analysis, we develop a model that accounts for the recreational value provided by urban green spaces. Our model includes three factors that affect this value: the quality of the green space; the distance to the green space, and characteristics of the affected individuals.

As access to green space close to homes is especially important for groups of individuals such as children, older persons, and individuals with disabilities for whom moving over longer distances is more difficult, we allow for a varying recreational value between individuals. These three groups are also mentioned in the UN's Global Sustainability Goals, Agenda 2030. For example, Goal 11.7 is to provide universal access to safe, inclusive, and accessible green and public spaces by 2030, particularly for women, children, older individuals and persons with disabilities. We illustrate our proposed method using the real-world example of a case in which green space was lost in Lund, Sweden. In our application we use benefit transfer and recreational values from other studies.

As different societal groups have different needs and reap different benefits from urban green spaces close to their homes, it is important to highlight the distributional effects of ecological compensation. The same argument is also valid for the general allocation of urban green spaces in urban planning. De la Barrera et al. (2016) apply a set of indicators (such as the total area of green space in relation to the population and the spatial distribution and accessibility of green space) that can be used to study distributional inequalities related to urban green space, but they do not include indicators for different categories of individuals. Based on British studies, Peroni et al. (2014) carry out a meta-analysis and estimate marginal value functions for recreational sites that are used to simulate the effects of urban growth due to changes in key urban parameters such as the area covered by settlements, the number of people living in such areas, and the amount of urban green space. The analysis is carried out at an aggregate level and does not study the effects on different categories of individuals.

Previous studies have also estimated the monetary value of urban green space. The approaches that have been used are mainly hedonic price studies (e.g., Panduro and Veie, 2013; Jim and Chen, 2010; Morancho, 2003) and stated preference studies based on the contingent valuation method or choice experiments (Bockarjova et al. 2020; Jim and Chen 2006; Bertram et al. 2017; Latinopoulos et al. 2016; Lo and Jim 2010). This information represents the individuals' preferences and willingness to pay for urban green space. However, the vast majority of these studies provide no information about how different individuals value urban green space. Exceptions are Jim and Chen (2006) and Lo and Jim (2010). These studies value urban green space in very large Asian cities, the characteristics of which are very different than those of Swedish and many other European cities. The valuation of urban green space from these studies is thus not ideal for benefit transfer (Brouwer 2000; Plummer 2009) to our model.

To measure the value of urban green space, we instead use the direct use value (Plottu and Plottu, 2007), based on information about different individuals' use of urban green space (parks) and information about the accessibility of urban green space. Although the tool that we develop in this study is applied to a case study in Sweden, the general modelling framework can be applied in other countries. Compared to previous models that have been used to evaluate loss of urban green space and compensation measures, this model contributes by adding welfare estimates for the valuation of urban green space by different groups of individuals. It can thus be seen as an extension of the approach used in de la Barrera et al. (2016), to which information is added about the well-being that urban green space provides different types of individuals.

The approach implies that we can compare the welfare benefits of compensation at various offset sites with the welfare losses at the impact site and study the distributional effects among different categories of individuals. The benefits of the model include that it is transparent and easy to use, and that it gives an approximate value of the welfare effects of compensation at various offset sites. Such information is of great value for city planners and the interested public.

This article is structured as follows: The modelling framework is outlined in the next section, and a case study follows in Section 3. The paper ends with a discussion in Section 4.

2. Proposed method for evaluating green space

In this section, we propose a method for evaluating changes in urban green space that allows us to compare losses of ecosystem services (recreation) in the impact area with benefits of ecosystem services (recreation) in the offset area. We start by describing our model intuitively, moving on to a more formal description in the latter part of this section. Figure 1 describes two types of ecological compensation that can be used when an impact area loses green space.

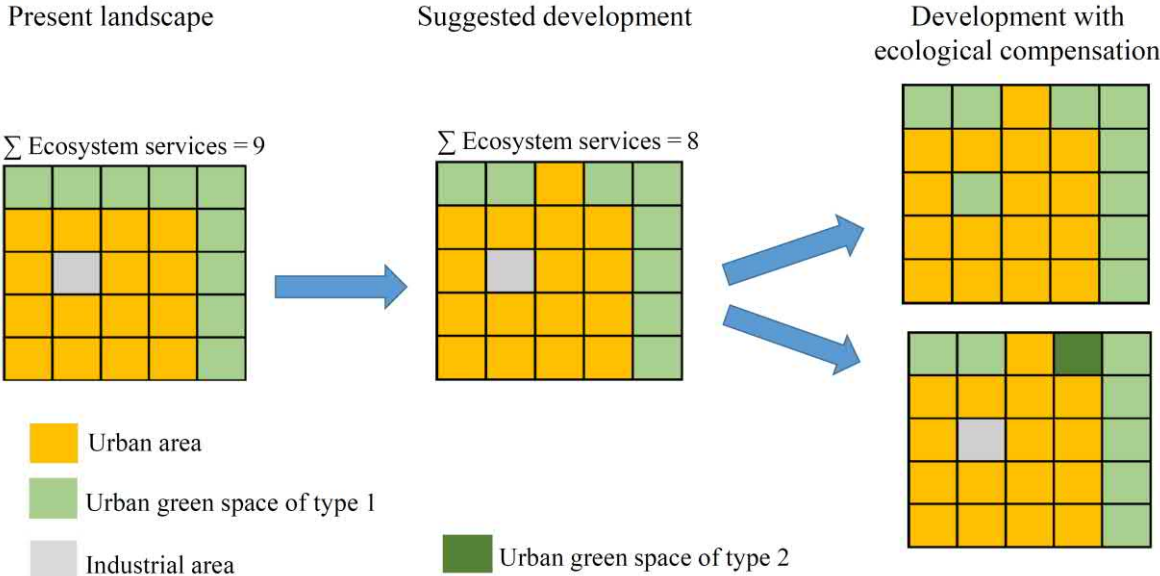


Figure 1. Utilization of a recreation area with compensation in different areas

The figure to the left shows the current situation, where there is an industrial area within the urban area and urban green space of type 1 surrounding the urban area. Each square of green space of type 1 gives a value of one to the inhabitants in the urban area (for the moment, the distribution of inhabitants is not important). The total value of green space in the situation at

hand here is nine. When part of the green space transforms to an urban area, the total value of green space is reduced to eight. To compensate for the loss of urban green space, there are at least two different alternatives:

- One alternative is to move the compensation to another part of the urban area and transform the industrial space to green space of type 1. The total utility of urban green space after compensation is nine in this case as well.
- Another alternative is to compensate as close as possible to the impact area and to increase the quality/value of the urban green space from type 1 to type 2 (the dark green area in the top figure to the right in Figure 1). Assuming that the utility or well-being of a green space of type 2 is twice as high as the utility of green space of type 1, this type of compensation will also result in a total utility of nine after the transformation of the green space.

From a no net loss perspective, these kinds of compensations are possible if the distance to urban green space is irrelevant for the inhabitants' utility and well-being, and if all individuals get the same utility of urban green space. For parks in urban areas, studies show that the distance to urban green space strongly influences use of the park (see e.g. Ekkel and de Vries 2017, Panduro and Veie 2013, Miller 2001).

In order to take distance and characteristics of the inhabitants in consideration, we add information about the distribution of the individuals within the urban area.

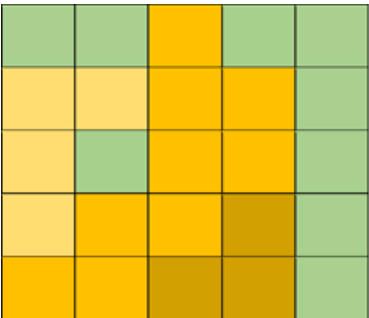


Figure 2. Example in which compensation is carried out in an urban area (the industrial area in Figure 1 turned into urban green space of type 1), with a heterogeneous population represented by the yellow-coloured squares.

The different shades of yellow in the squares in Figure 2 represent differences in demographics, e.g., differences in the number of adults and children and the individuals' education level, age, or income, but they can also represent the density of the population within the square. Light yellow could represent e.g. a residential area with few inhabitants, dark yellow may denote an area with many inhabitants, and the intermediate yellow colour could show an area with an average number of inhabitants. With information about the geographical location of the squares the distance to green space can also be calculated. With these extensions, the aggregate utility of urban green space will differ depending on which yellow square we are studying. Compared to Figure 1, the total utility of urban green space in Figure 2 may be greater or lesser than nine.

The provision of ecosystem services (e.g. recreation), the subsequent benefits and the well-being of humans is supported by ecosystem processes that are themselves driven by biological diversity (Balvanera et al., 2006). The no net loss principle should thus also be applied to biodiversity. Griffiths et al., (2018) and Moilanen and Kotiahe (2018) suggest however that the analysis for compensation of ecosystem services and biodiversity can be performed separately.

More formally, we can describe the total utility of urban green space as a function of the quality of the green space (q), the distance to the green space (d), and the characteristics of individuals (k). The total utility U derived from green space for all affected individuals i , ($i = 1, \dots, N$), can then be expressed as:

$$U(i, k, d, q) = \sum_{i=1}^N u_i(k, d, q). \quad (1)$$

Applying the no net loss criteria to the aggregate utility from urban green space means that the gain in utility from improved urban green space in the offset area should be at least as large as the loss of utility in the impact area. Denoting the gain in utility from urban green space in the offset area at u^O and the loss in utility in the impact area as $-u^I$, the no net loss criteria in utility terms can be expressed as

$$\sum_{i=1}^M u_i^O(q, d, k) - \sum_{i=1}^N u_i^I(q, d, k) \geq 0, \quad (2)$$

where M and N denote the number of individuals that are affected by the green space in the impact- and offset area, respectively. These numbers (individuals) can be the same or differ between the offset- and impact area, depending on where the offset area is located.

2.1 Measuring utilities

The utility from urban green space derives from the eco-system services that the space provides. Recreation possibilities could have direct user values as well as option values. Direct user values are the utility or well-being that one derives from visiting the green space (if one lives close to the green space, this may also be the utility derived from the view of the green space). Option values are values related to the possibility to use the green space in the future, e.g., if an individual that is currently not using a green space plans to use it upon retirement from the labour market.

A monetary valuation of urban green space, e.g. willingness to pay (WTP) for urban green space, could be used to calculate the utility expression in (1) by summing the N individuals' WTP to preserve the urban green space in the impact area. However, the recreational value or the utility of urban green space can also be represented by the direct use value, which can be measured by how often individuals use urban green space, e.g. visit a park for recreational purposes. Schipperijn et al. (2010) study how different groups of individuals use urban green space by estimating the probability that different individuals visit a park at least once a week. The study is based on a representative sample, $n = 11238$, of the Danish population.

Results in the Fredman and Hedblom (2015) and Schipperijn et al. (2010) show that recreational patterns and outdoor life are similar in Sweden and Denmark. Findings by Sang et al. (2016), who study Swedish citizens' experience and use of urban green space, are also in line with the results in Schipperijn et al. (2010). We thus consider Schipperijn et al. (2010) a reliable study for benefit transfer and our calculations of recreational utility of urban green space in Sweden. In the calculations of our utility expression, we see utility as a cardinal utility that can be represented numerically. Our aggregate utility expressions in equations (1) and (2) can thus be represented by a figure or utility index, based on how often individuals with different characteristics visit urban parks. Based on the specification of the econometric model and reported results in terms of the odds ratios for visiting a park in Schipperijn et al. (2010), we adjust our utility expression in equation (1), so that we can use the odds ratios as weights in the utility index. In the regression model, a woman aged 46-65 years with a university education is used as the reference individual (with an odds ratio of one).

Table 1: Odds ratios for an individual to visit the green area at least once a week.

Variable	Odds ratio	95% Confidence interval	Weight in utility index
Men			
16-24 years	0.43	0.32-0.58	0.43
25-44 years	0.52	0.44-0.60	0.52
45-64 years	0.79	0.68-0.92	0.79
65-79 years	1.03	0.84-1.26	1
80 years and older	0.53	0.37-0.77	1
Women			
16-24 years	0.80	0.62-1.04	1
25-44 years	0.66	0.57-0.78	0.66
45-64 years	1		1
65-79 years	0.92	0.75-1.13	1
80 years and older	0.40	0.28-0.55	1
Education			
Secondary school	0.81	0.70-0.95	0.81
Upper secondary school	0.85	0.76-0.95	0.85
University	1		
Distance from the green area			
Less than 300 m	3.26	2.96-3.60	
300 m – 1 km	1		
More than 1 km	0.41	0.34-0.49	

Source: Schipperijn et al. (2010), *Factors influencing the use of green space: Results from a Danish national representative survey, Landscape and Urban Planning*.

Table 1 thus gives us information about the relative utility for individuals with different characteristics. The odds ratio of 0.52 for men aged 25-44 years implies that the odds that these men will visit a park are 48 percent lower than for the reference individual. If we normalize the utility of visiting a park to 1 for the reference individual, we can interpret the results to express that men aged 25-44 years have 52 percent of the utility of the reference person of a park (urban green space). We can then use the odds ratios for visiting a park as utility weights in the calculations of the aggregate utility in equation (1).

After normalization of the utility expression in equation (1), the utility expression can be written as

$$W(i, k, d, q) = \sum_{i=1}^N \frac{u_i(k, d, q)}{u_{i=ref}(k, d, q)} = \sum_{i=1}^N \frac{u_i(k)}{u_{i=ref}(k)} * \frac{u_i(d)}{u_{i=ref}(d)} * \frac{u_i(q)}{u_{i=ref}(q)} = \sum_{i=1}^N w_i(k) * w_i(d) * w_i(q) \quad (3)$$

where $u_{i=ref}(\cdot)$ is the utility of the reference individual of a green space (park), and $w_i(k)$ is the odds ratio for individual i with characteristics k to visit a park compared to the reference

individual; $w_i(d)$ is the odds ratio for visiting a park depending on the distance to the park and $w_i(q)$ is the odds ratio for visiting a park depending on the quality of the park.

If all individuals have the same probability (utility) of visiting a green space, the normalized utility $w_i(.)$ would be 1 for all individuals. Therefore, if the estimated odds ratio for a specific group of individuals in Table 1 is significantly different from 1 (the value for the reference individual), we use the estimated odds ratio in the calculations of the utility index; otherwise, we use the value 1 for the utility weights.

Schipperijn et.al. (2010) do not include children in their study, but we attach a value of 1 for the utility weights to children, since children are a prioritized group in the Agenda 2030 goals. This is the same value as that for individuals aged 65-79 years and women with an age of 16-24 and 45-79 years in the utility index, and it is also the highest value that is used in the index for individuals with different characteristics; see Table 1. Since older persons are also a prioritized group in the Agenda 2030 goals, we attach a value of 1 to the utility weight for individuals 80+ as well, although the odds ratio for this group is significantly lower than 1. In a later sensitivity analysis, we will examine the sensitivity of our results with variations in the utility weights.

Previous studies show that the value of urban parks declines sharply when the distance to the park increases (see e.g. Ekkel and de Vries 2017, Panduro and Veie 2013, Jim and Chen 2010, Miller 2001). The proximity measure in Schipperijn et al. (2010) is thus not sufficiently detailed for use in the calculations of recreational utility of urban parks. For regular visits by foot to urban green space, studies have found that the distance to the green space should not exceed 300m (Boverket 2007). This distance is also of great importance for children and older persons, whose possibilities to move longer distances are frequently more limited (Fredman and Hedblom 2015). 100-300m was found to be a threshold distance after which use falls rapidly, although larger distances are not irrelevant (Ekkel and de Vries 2017). For that reason, we complement the results in Schipperijn et al. (2010) with the findings in Panduro and Veie (2013) that estimate the impact of distance to urban parks on house prices in Denmark, and thereby give a clear indication of individuals' preferences for proximity to parks. Information from Panduro and Veie (2013) is available for a distance up to 600m; this is divided into different distance intervals. Adjusting Schipperijn et al.'s (2010) odds ratios with the information in Panduro and Veie (2013), for the distance intervals that we use in the case study, gives the

following utility weights: 3 for 0-250m, 1 for 250-500m, and 0.21 for 500-750m. See Appendix A for the calculations of the distance utility weights.

3. Case study – the loss and gain of green space in Lund, Sweden

In this section, we apply our proposed method to the loss of a green space in Lund, Sweden. Lund is a city in southern Sweden and the central locality of Lund municipality. The municipality had 125 000 inhabitants in 2019 (Statistics Sweden, 2020), and it is characterized by its university and its research and development-intensive industries. The number of inhabitants increased from 109 000 to 125 000 between 2009 and 2019 (Statistics Sweden, 2020), making Lund one of Sweden's fastest growing municipalities. This also means that the development and building on green space in the city has been significant in recent years (Lund Municipality, 2015). The municipality is expected to continue to grow in the future (Lund Municipality, 2018).

3.1 Data

Data was obtained from three different registers from Statistics Sweden: the integrated database for labour market research (LISA); the population register (RTB), and the geographic database. Data on the age, sex, marital status, disposable income, the number of children under the age of 20 and the number of years of schooling is available for all individuals aged 15 years or older. In addition, the area of residence is defined as a 250×250-meter square. All data is for 2017.

3.2 The loss of green space and possible compensation

Our case study looks closer at a green space in the eastern part of the city known as Vipeholmsparken that was lost in the late 2010s. Lund Municipality suggested building 550 apartments in 24 apartment blocks according to the local plan for the area drawn in June 2015 (Figure 3). When the planning was initialized, the area was characterized by a pre-school and an upper secondary school in a park-like environment (Lund Municipality, 2015). The plan stated that much of the park-like character of the area would be lost, and that especially, the changes in the northern part of the planning area would be substantial. Pedestrians and cyclists use a pathway located there for recreational purposes (Lund Municipality, 2015). The map of the area from the planning documents is presented in Figure 3.



Figure 3: The Vipeholm area, where housing development commenced in 2015. Source: Lund Municipality, 2015.

The most recent planning documents from the city of Lund states that ecological compensation can be used when green space is exploited (Lund Municipality, 2018). We thus suggest two different alternatives for the developer to compensate the loss of green space in the Vipeholm area. The first alternative is for the developer to compensate on the outskirts of the city, relatively close to Vipeholmsparken, and to construct a green space of equal size and quality there. The second alternative is to compensate in the western part of the city, where new housing development is being planned for an area called Västerbro. We assume that part of the development of this area – which is currently used for industrial- and commercial purposes – could be used to compensate the loss of Vipeholmsparken. Whereas the agricultural land on the outskirts of the city is relatively close to the impact area, at a distance of approximately 1500m,

the Västerbro area is located further away, on the other side of the city centre. The map in Figure 4 shows the locations of the impact- and offset areas in Lund.



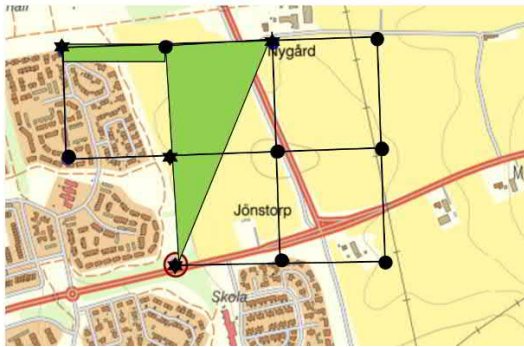
Figure 4: Map of Lund showing the site where green space is lost (red triangle) and two possible offset sites (green triangles).

Using information from municipality planning documents, we chose four coordinates assumed to be entry points to Vipeholmsparken. As the green space is not enclosed, we assume that entry was possible where pathways led to the green space. In addition, we chose one set of coordinates along the pedestrian pathway on the northern side of the area since this part of Vipeholmsparken was indicated as of special importance for the appearance of the green space.

For the sake of simplicity, we assume that the green spaces in the offset areas are identical to the lost green space in size and quality (in the later sensitivity analysis, we forgo this assumption), i.e., we do not consider changes in biodiversity, instead we assume that the offset area contains the same amount of biodiversity per surface unit as the impact area, although the mix of biodiversity can vary. The assumptions about equal quality of the urban green spaces may be supported by the findings in e.g. (Bronnmann et al., 2020), who find that when individuals were confronted with different alternatives for their closest green space, they showed a clear preference for the status quo-type park.

Figures 5a and 5b show schematically how green space has been placed in the two alternative offset areas, on the outskirts of the city and in the city centre, in Västerbro. The grids show examples of 250×250m squares located around the green spaces.

(a)



(b)

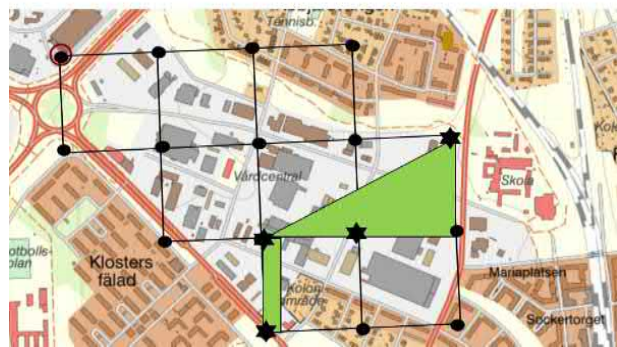


Figure 5a and 5b: Compensating the loss of green space in two different offset areas, on the outskirts of the city and in the city centre, in Västerbro. The 250×250 -squares in Figure 5b approximately indicate the area to be developed, and the green space is placed in the east part of the area.

Since the municipality is planning for housing development in the Västerbro area, we need to make assumptions about the individuals who will eventually live in the area. Planning documents describe the new development as having a “mixed city character” and as being a “continuation of the city” (Lund Municipality, 2018). Our assumptions about the population structure in the new Västerbro area are thus based on the population structure in nearby areas. We have chosen three squares in our data that are situated between the current Västerbro area and the city centre, and we use these to approximate the population structure (see Appendix B for details).

In order to find affected individuals at the impact- and offset areas, we use the cut-off points based on information from Panduro and Veie (2013) described above. As mentioned earlier, green space is assumed to provide utility only when individuals reside less than 750m from an entry point to the green space. We thus calculate the distance from the entrance points of our green spaces to the residence of all individuals in Lund Municipality by using the coordinates of the entrance points to the green spaces. As we do not have the exact coordinates for individuals’ residences, we use the coordinates at the centre of each 250×250 -square to define the point of residence for the individuals in each square. Table 2 shows a description of the affected areas.

Table 2: Descriptions of affected areas in our example

Impact area	Offset areas	
Area around Vipeholm	Area on the outskirts of the city	Västerbro
Inhabitants lose green space in this area	Inhabitants gain green space in this area. New green space is constructed on agricultural land.	Inhabitants gain green space in this area. New green space is constructed as part of the construction of residential areas.

3.3 Descriptive statistics of the affected areas

From the descriptive statistics in Table 3, it is clear that there are more individuals in the Västerbro area than in the other two areas further from the city centre. The lattermost have the smallest number of residents; this is to be expected, since there are residential areas only on the west side of the offset area/agricultural land. The area around Vipeholmsparken has a greater number of individuals than the area on the outskirts of the city, but there are actually fewer individuals living 250-500m from the green space. The two areas surrounding the impact and offset area on the outskirts of the city overlap to some extent.

Looking at the share of individuals in different age categories in the different areas, there are some notable differences. The area around Vipeholm has relatively many young adults (16-24 year-olds) and more individuals older than 65 years. In particular, Vipeholm is the area with the largest share of individuals above 80 years of age (8 %). The area on the outskirts of the city, on the other hand, has a greater number of children and individuals aged 45-64 than the other areas. In fact, one quarter of the population in the area is younger than 16. Notably, when compared to the other two areas, the number of young adults (aged 25-44 years) is higher in the Västerbro area. The share of women in the population is similar in all three areas.

Table 3: Number of individuals and characteristics of individuals in the affected areas

	Area around Vipeholm	Area on the outskirts of the city	Västerbro
<i>Distance</i>	Number of individuals in affected areas		
0-250 m	1725	1141	4700
250-500 m	1973	2702	4598
500-750 m	4510	2594	7256
Total	8208	6436	16553
Share of individuals in different age groups			
Children (0-15 years)	15%	25%	12%
16-24 years	19%	10%	19%
25-44 years	23%	28%	34%
45-64 years	21%	24%	21%
65-79 years	13%	10%	11%
80 years and older	8%	3%	3%
Share of women in adult population (>15 years)			
	52%	52%	52%
Share with primary or secondary education			
	39%	38%	38%
Share with tertiary education			
	61%	58%	62%
Average disposable income (SEK)			
	286 400	266 000	223 900
Median disposable income (SEK)			
	191 800	239 400	NA

As shown in Table 3, there are some differences in incomes between the areas; individuals around Vipeholmsparken are wealthier on average than individuals in other areas. Individuals around the Västerbro area have lower average incomes than individuals around Vipeholmsparken and in the area on the outskirts of the city. However, the median disposable income suggests that income variation is greater in the area around Vipeholmsparken than in the area on the outskirts of the city. The level of education is very similar in the different areas. Around 40 percent of the population have primary or secondary education, and around 60 percent also have tertiary education. It is thus more likely that the differences in incomes are related to age rather than to education.

As mentioned, we will attach utilities of green space to individual characteristics. In our application, we will base this link on age and sex. It would, however, also be possible to use other individual characteristics to define the utility of urban green space, on the condition that data is available on utilities. While our utility calculations will not use income or education

variables, we will use these variables to discuss distributional effects for the different compensation alternatives.

3.4 Results

Table 4 shows the utility of green space, the number of inhabitants, and the utility per individual for the areas surrounding the impact and offset areas. The total utility of green space (of the same size and characteristics) is greater in the area around Vipeholmsparken than in the area surrounding the offset area on the outskirts of the city. Achieving a no net loss of utility by compensating with a green space of equal size and quality on the outskirts of the city is thus not possible. The net loss is 1216 units. However, the results show that the total utility of green space is greater in the Västerbro area than in the impact area around Vipeholmsparken. In our simple example, it is thus possible to compensate the loss of green space around Vipeholmsparken by constructing green space in Västerbro. The gain is 8956 units.

Table 4. Utility, population, utility per individual and net utility in three scenarios for the area on the outskirts of the city, Västerbro (offset area) and for the area around Vipeholmsparken (impact area).

Distance	Vipeholmsparken				Green space on the outskirts of the city			
	Utility	Population	Percent of pop.	Utility per capita	Utility	Population	Percent of pop.	Utility per capita
0-250 m	4282	1725	21	2.48	2858	1141	18	2.51
250-500 m	1723	1973	24	0.87	2236	2702	42	0.83
500-750 m	762	4510	55	0.17	457	2594	40	0.18
Total	6767	8208	100	0.82	5551	6436	100	0.86
-1216								
Västerbro with houses and green space								
Distance	Utility	Population	Percent of pop.	Utility per capita				
0-250 m	10863	4700	28	2.31				
250-500 m	3641	4598	28	0.79				
500-750 m	1219	7256	44	0.17				
Total	15723	16553	100	0.95				
Net utility	8956							

Utility per capita is higher on average in the offset area on the outskirts of the city than in the area around the impact area Vipeholmsparken (0.86 compared to 0.82). The main reason for this is that a larger share of the population resides relatively close to the suggested offset area on the outskirts of the city (in particular, there are not as many individuals residing within 500-

750m of the green space on the outskirts of the city). The greatest average utility per capita is found in the offset area Västerbro (0.95). According to our estimates, when Västerbro is developed, the utility of green space will increase substantially, as there are many individuals within 250m from the projected green space.

The characteristics (age and gender) of the population do not play a major role for this outcome; instead, the size of the population that can access the green space is decisive. Although a typical inhabitant around Vipeholmsparken has greater utility from the green space, the number of affected individuals close to the green space is the dominating factor explaining why the utility of green space in the Västerbro area is greater than the utility of green space in the Vipeholm area.

Regarding the distribution of welfare effects, clear winners and losers can be discerned in the spatial dimension. The winners will be those who live in the area surrounding the offset area, and the losers will be those who live in the area surrounding of the impact area. If we look at the distributional effects in the income dimension, we see that on average, low-income individuals will benefit if the impact area is exploited and the compensation is carried out in any of the offset areas. However, there is a much more uneven income distribution in the impact area, with many pensioners (individuals aged 65+) among the residents, so when one considers both the mean and median income, the distributional effects in the income dimensions become less clear. The individuals' education level is relatively evenly distributed between the different areas, so the distributional effects are small in the educational dimension. The most prominent distributional effects are instead found in the age dimension, where older persons will lose out if the impact area is exploited, and children (below an age of 15 years) will win if the compensation is carried out on the outskirts of the city.

3.5 Sensitivity analysis

In this section, we perform a sensitivity analysis to investigate the robustness of the results. As the most important factors affecting the results are the number of inhabitants living close to the investigated green space and the utility weight that depends on the distance from green space, we will first study how changes in the utility weights affect the results. Table 5 shows three alternative assumptions (i.-iii.) about the distance-dependent utility. The first alternative (i.) assumes that individuals residing further than 250m from the green space have greater utility than what was assumed in the original analysis. The second alternative (ii.) assumes that all individuals within the 750m-limit have the same distance-dependent utility, and that this utility

is set to one. Finally, the third alternative (iii.) assumes that individuals living close to the green space have significantly greater utility, and that the utility is only experienced within 500m of the green space.

Table 5: Sensitivity analysis of total utility with different assumptions about the value derived at different distances from the park.

	Original	i.	ii.	iii.
<i>Distance</i>				
0-250 m	3	3	1	4
250-500 m	1	2	1	1
500-750 m	0,21	1	1	0
<i>Total utility</i>				
Vipeholmsparken	6767	11357	6779	7432
Green space on the outskirts of the city	5551	9506	5364	6047
Västerbro with houses and green space	15723	23951	13067	18125

The results shown at the bottom of Table 5 imply that changing the distance-dependent utilities does not significantly alter the results. The aggregate utilities are smaller around the offset area on the outskirts of city regardless of the choice of distance-dependent utilities in Table 5. The aggregate utilities are still greater from the green space in the offset area in Västerbro than from the green space in the impact area around Vipeholmsparken.

Although the utility that depends on the characteristics of the individuals is relatively modest compared to the distance-dependent utilities, one should remember that we have assumed a value of 1 for children and older individuals in our index. Even if this is the highest weight attached to an individual characteristic in our index, it could still be too low (or too high) compared to the true value. In Table 6, we show how the results change if we change the utility for children (aged 0-15 years) and individuals aged 65 or above. We work with three different alternatives: first, we double the utility for children and individuals aged 65 and above, and second, we leave the utility for individuals aged 65 and above unchanged while the utility for children doubles, and finally, we allow the utility for children to increase to 4 whilst leaving the utility for individuals older than 65 unchanged.

Table 6: Utility when using the assumption that children and individuals aged 65 and above have higher utility than in the original analysis.

	Original	Utility for children and 65+ = 2 ^a	Utility for children = 2 ^a	Utility for children = 4 ^a
Vipeholmsparken	6767	9834	8105	10782
Green space on the outskirts of the city	5551	8112	7269	10705
Västerbro	15723	17946	17714	21694

Note: ^a instead of a utility of 1 in the original analysis.

Starting with the first alternative, we see that increasing the utility for older individuals and children does not change the relationship between different areas substantially. Just as in the original analysis, constructing a park of equal size and quality in the Västerbro area is more beneficial regardless of the chosen scenario. Although there are relatively more individuals older than 65 and children in the Vipeholm area and the area on the outskirts of the city (see Table 3), the number of individuals in the areas remains the most important factor for total utility. The number of children is significantly greater around the green space on the outskirts of the city, and the second analysis shows that when only the utility of children is doubled, the utility is relatively great in this area. However, it is not great enough to change the order of total utility between areas. When the utility for children is assumed to be four (the final alternative), more children on the outskirts of the city result in a total utility that is almost as high as in the area around Vipeholmsparken. Thus, if children are assigned significantly higher utility from green space than adults, it will be interesting to compensate for the loss of green space in areas with many children in our example. On the other hand, our analysis shows that a green space of equal size and quality has greater utility in the area around Västerbro, as the number of inhabitants, children and others, is higher there.

Finally, we forgo the assumption that the green spaces in our areas are of the same size (in the analysis, we still assume that the quality of one hectare of green space is the same in all three areas and that the same number of individuals lives in the areas around the offset- and impact areas). For this analysis, we use information from Peroni et al. (2014), who carry out a meta-analysis and estimate a marginal value function for urban green space. The marginal value function is a function of the distance to green space, the size of the green space, the size of the population in the city, and income. As expected, the utility or marginal value diminishes with size, i.e., each hectare of additional green space is worth less from the perspective of the individual. In particular, we study how large the green space has to be at the different offset areas in order to offset the loss of utility from Vipeholmsparken exactly; that is, we keep the

utility of urban green space in the city constant at the original level of 6767 and adjust the size of the offset areas to reach this utility.

Table 7: How much green space is necessary to achieve the same utility in offset areas as in the area around the impact area Vipeholmsparken?

	Hectares
Vipeholmsparken (impact area)	2.50
Green space on the outskirts of the city (offset area)	3.72
Västerbro (offset area)	0.46

Table 7 shows that to achieve the same utility level on the outskirts of the city, the green space must be 1.22 hectares, or 49 percent larger than the green space in the area around Vipeholmsparken. However, it is possible to compensate for the loss of the green space by constructing a much smaller green space in Västerbro. To arrive at the same utility from green space in Västerbro as in the area around Vipeholmsparken, the area needs only to be one fifth of the lost green space. This is a considerably smaller green space. A word of caution is warranted here though: we have not taken other green spaces in the areas into consideration, and crowding might make a smaller green space less valuable to the inhabitants.

4. Discussion

In this study, we have developed a transparent, simple and an easy-to-use tool that can be utilized to evaluate the welfare effects of ecological compensation in various offset areas. The model can also be used to evaluate the so-called no net loss criterium, which entails that the change in welfare should be non-negative after compensation. The welfare effect and ecosystem service that we include in our analysis are recreational values from urban green space. The model is applicable to urban green space such as parks and other green spaces used for recreation. Green spaces that contain endangered species must be protected and are not considered possible impact areas. In the case study, we have assumed that there is no net loss of biodiversity. The loss of biodiversity at the impact area is assumed to be replaced with a new mix of biodiversity at the offset area that gives the same recreational values per surface unit (e.g. hectare). This assumption has been made to facilitate the analysis. A separate evaluation should be done for biodiversity to support the analysis of the effects on ecosystem services.

We carried out a case study to demonstrate how the model can be used. The results from the case study indicate that the differences in valuation of urban green space among different groups of individuals have a fairly small impact on the aggregate welfare measure: the most important

factor for the aggregate welfare measure is the number of individuals that have access to the recreational area; i.e., the number of individuals that live close to the urban green space.

To measure the welfare or well-being of urban green space, we used the direct use value, represented by the likelihood that various groups of individuals visit urban green spaces. Monetary valuation studies have shown that individuals with higher incomes tend to have a higher willingness to pay for urban green spaces. On the other hand, studies also show that individuals with lower incomes use urban green spaces in close proximity to their homes to a greater extent than individuals with higher incomes. One explanation is that they can't afford to visit green spaces at a greater distance from their homes (Peroni et al. 2014). The lower WTP for urban green space may thus suggest that the budget constraint is a limiting factor for low-income groups when they state their WTP for urban green space. From that perspective, the use value tends to be a more neutral welfare measure with respect to income. That may also be an advantage in the study of the distributional effects of ecological compensation and the spatial distribution of green space in general. An analysis in which the welfare measure is based on WTP may favour areas with higher incomes.

An additional advantage of the use value is that it is easier to measure the use of green space for children than it is to measure children's WTP for urban green space. In monetary valuation studies, the value of urban green space is reflected in the parents' WTP. The study that was used for benefit transfer and the calculations of the use value of urban green space did not include children aged less than 15 years. To resolve this, we imposed a value corresponding to the highest use value that was revealed in the survey. The true value may be greater or less than the imposed value. We suggest that future studies aiming to estimate the use of urban green space also study the use of children below 15 years. For some age groups below 15 years, visits to urban green spaces in close proximity to the home may be high. It is thus also of interest to study the visiting frequency in more detail, and to not limit the analysis to visits to urban green space that are done at least once a week, as was done in the study that we used for benefit transfers.

In future valuation or use studies of urban green space, it would also be of interest to study how different sizes of green space, additional green spaces (parks) in close proximity to the residential area and crowding of green space affect the use and perceived value of urban green space. As our case study shows, information about the marginal value of size and crowding would have been useful with regard to the offset area in the city centre. In this study, the developer is assumed to compensate directly through an in-situ project. The cost for ecological

compensation thus depends on the price for land and the size of the offset area. If the developer was given the option to compensate in the central part of the city or in the outskirts of the city, s/he may choose to compensate in the outskirts of the city, although s/he would have to compensate with a larger area than the offset site the in central part of the city, as the option would be cost efficient.

We have assumed that we can compensate the loss of green space in one part of the municipality when another part is developed. It is important, however, that additional green space is added in the municipality: one of the principles of ecological compensation is the additionality principle, meaning that lost green space must not be replaced by green space whose construction is planned regardless of whether green space is lost. For our offset area at Västerbro the plan is to build a new residential area. It is unlikely that a residential area will be built without green space, so from a compensation perspective it is important that the offset area adds green space on top of what was already planned prior to compensation for the impact area becoming a reality. The smaller size of the offset area in Table 7, which achieves the same utility as the utility of the impact area, may thus be a more realistic assumption regarding the size of an area that fulfils the additionality requirement.

Another extension that may be of interest in the calculations of the aggregate welfare index is inclusion of all green spaces in close proximity to the individuals' residences (in our case study, within 750m). In our case study, we studied the effects of losing a park at the impact area and the benefits of adding green space for recreational purposes at the offset area. This was done to evaluate how our suggested welfare index was affected by differences between different individuals' valuations of urban green space and to study how the welfare measure was affected by other factors such as population density and proximity to the green area. For policy makers, knowing whether or not there are other green spaces in close proximity to the individuals' home is also of interest. If there are other green spaces in close proximity to the individuals' home, the loss or gain of green space is probably smaller than if there is only one green area in the vicinity, similar to individuals' marginal valuation of a larger sized green area. As it may take some time before the offset area can provide the ecosystem services (e.g., it takes time for a tree to grow), the benefit and losses of welfare may also be discounted. In our case studies and specification of the no net loss in equation 2, we implicitly assumed that the gains at the offset area would take place at the same time as the losses at the impact area.

We have also shown that ecological compensation will have distributional effects unless the offset area is very close to the impact area. Some individuals will gain recreational utility, whilst

others will lose it. The decision on how to distribute urban green space is a question for the policy makers, and the tool that we have presented here can be used in that process. The transparency of the tool, which aims to capture the well-being of all individuals, entails that groups of individuals that may find it difficult to make their voices heard in the decision process (such as groups of individuals with a lower education and children) also are represented in the decision material.

It should also be noted that individuals might not value the loss of green space the same way as the gain of green space. It is possible that the reference level, i.e., how much green space is expected in the area, for those who lose differs from that of those who gain. That utilities from loss differ from utilities from a gain was first suggested by Kahneman and Tversky (1979), and losses are often found to be valued more highly than gains of the same magnitude in the empirical literature (Knetsch et al., 2012). For example, in the area around Vipeholmsparken, the reference level is a state in which there is a green space in the area. The value of this loss per individual might be greater than the value of the gain of a green space per individual on the outskirts of the city or in Västerbro, where the reference level of green space is a situation in which there is no added green space.

Appendix A

Calculations of odds ratio for distance based on estimated coefficients (marginal values) in Panduro and Veie (2013). Odds ratio = Odds of group 1/Odds of group 2.

Odds of Group 1: the odds that people living 0-250 m from the park visit the park.

Odds of Group 2: the odds that people living 250-500 m from the park visit the park.

The assumption behind the hedonic price study is that the value of a park (visits to a park) is reflected in house prices. In the econometric model, these effects are quantified for houses/apartments at various distances from the park.

The odds ratio for visiting a park at various distances from the park can thus be written as:

$$\frac{\text{marginal value of a park for a house 0 – 250 m from the park}}{\text{marginal value of a park for a house 250 – 500 m from the park}}$$

where the marginal values reflect how often the houseowner (or the members of the household) visits the park. As the estimation results in Schipperijn et al. (2010) normalize the odds ratio to 1 for houses 300-1000m from a park, we have chosen to normalize our odds ratio with respect to houses 300-500m from the park in order to make our results comparable to those of Schipperijn et al., as it correspond to the distances that we use in our analysis. We use the percentage change in house prices associated with a 100m decline in distance to a park as the marginal value of a park.

The results in Table A are from Panduro and Veie (2013) and are based on a regression model with house- or apartment prices as the dependent variable and the distance from a park as explanatory variables.

Table A. Green space appreciation index. Percentage change in price associated with a 100m decline in distance.

Houses	Distance to park						Sum of coefficients
	100m	200m	300m	400m	500m	600m	
Estimated coefficient	2.7	2.3	1.8	1.4	0.9	0.5	9.6
Share of value	0.281	0.240	0.188	0.146	0.094	0.052	
Accumulated share 100-300 m			0.708				
Accumulated share 400-500 m					0.240		
Apartments							
Estimated coefficient	2.1	1.7	1.4	1	0.7	0.3	7.2
Share of value	0.292	0.236	0.194	0.139	0.0972	0.042	
Accumulated share 100-300 m			0.722				
Accumulated share 400-500 m					0.236		

Note: Share of value = estimated coefficient (X metres)/sum of coefficients

Houses

Normalise with the value at the distance 300-500m

“Odds ratio” < 300m: $0.708/0.24 = 2.96$

“Odds ratio” 300-500 m: $0.240/0.24 = 1$

“Odds ratio” > 500m: $0.050/0.24 = 0.21$

Apartments

Normalise with the value at the distance 300-500m

“Odds ratio” < 300m: $0.722/0.236 = 3.06$

“Odds ratio” 300-500m: $0.236/0.236 = 1$

“Odds ratio” > 500m: $0.050/0.236 = 0.21$

For the “odds ratio” < 300m, we use the value 3 in the calculations of our utility index.

Appendix B

Since the proposed changes to green space are accompanied by changes in the housing development in Västerbro, we define a scenario for the future population structure in Västerbro. We will have to make assumptions about the number of individuals that will eventually live in the area, as well about the characteristics of these individuals. Our assumptions about the population structure in the new Västerbro area are based on the population structure in areas close to the area that is to be developed. In planning documents, the new development is described as having a “mixed city character” and as being a “continuation of the city” (Lund Municipality, 2018). To approximate the population structure of the new area, we have chosen three 250×250 squares in our data that are situated between the Västerbro area and the city centre (see Figure A1).

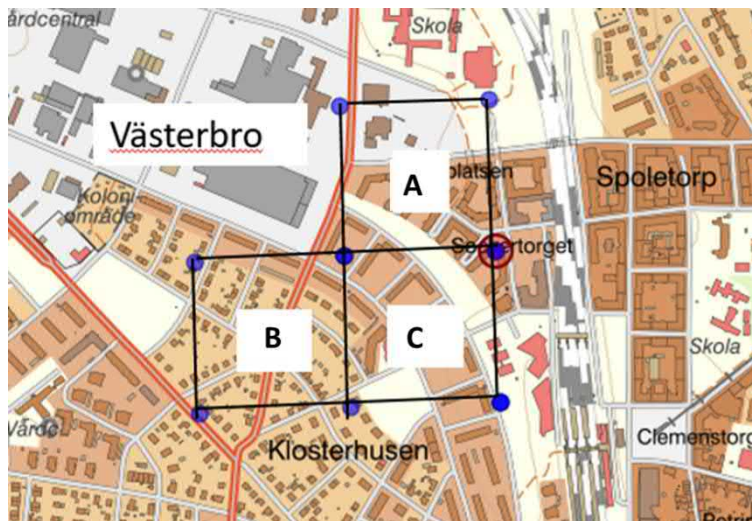


Figure A1: Three squares are used to approximate the population structure of the future Västerbro area.

Figure A1 shows the three squares. Square A has a city-like structure and was built quite recently. The four Västerbro squares closest to the city centre will be similar square A in Scenario 1. Since the new area is described as being of “mixed city character”, we will also use squares B and C to allow for the new area to include blocks of apartments, as well as villas. Two squares further from the city centre will be similar to square B, and two squares will be similar to square C in the Västerbro scenario.

References

- Balvanera, P., Pfisterer, A.B., Buchmann, N., He, S-J., Nakashizuka, T., Raffaelli, D. and Schmid, B. (2006). Quantifying the evidence for biodiversity effects on ecosystem functioning and services. *Ecology Letters*, 9, 1146–1156.
- BBOP (Business and Biodiversity Offsets Programme), 2009. *Business, Biodiversity Offsets and BBOP: An Overview*. BBOP, Washington, D.C
- BBOP (Business and Biodiversity Offsets Programme), 2012. *Standard on Biodiversity Offsets*. BBOP, Washington D.C., USA.
- Bertram, C., Meyerhoff, J., Rehdanz, K., Wüstemann, H., 2017. Differences in the recreational value of urban parks between weekdays and weekends: A discrete choice analysis, *Landscape and Urban Planning*, 159, 5–14.
- Bronnmann, J., Liebelt, V., Marder, F., Meya, J. and Quaas, M.F. (2020). The Value of Naturalness of Urban Green Spaces: Evidence from a Discrete Choice Experiment. Available at SSRN: <http://dx.doi.org/10.2139/ssrn.3721942>
- Brouwer, R. (2000). Environmental value transfer: State of the art and future prospects. *Ecological Economics*, 32, 137–152.
- Bockarjova, M., Botzen, W. and Koetse, M.J. (2020). Economic valuation of green and blue nature in cities: A meta-analysis, *Ecological Economics*, 169, 106480.
- Boverket. (2007). Bostadsnära natur - inspiration & vägledning. The Swedish National Board of Housing, Building and Planning. . Downloaded [2021-08-12] In Swedish
- Bull, J. W., Gordon, A., Watson, J. E. M., & Maron, M. (2016). Seeking convergence on the key concepts in 'no net loss' policy. *Journal of Applied Ecology*, 53(6), 1686-1693. doi:10.1111/1365-2664.12726
- Bull, J.W., Hardy, M.J., Moilanen, A., Gordon, A., (2015). Categories of flexibility in biodiversity offsetting, and their implications for conservation. *Biol. Conserv.* 192, 522–532.
- Bull, J. W., and Strange, N. (2018) The global extent of biodiversity offset implementation under no net loss policies. *Nature Sustainability*, 1(12), 790-798. <https://doi.org/10.1038/s41893-018-0176-z>
- Colding, J. (2011) The Role of Ecosystem Services in Contemporary Urban Planning. In *Urban Ecology*; Oxford University Press: Oxford, UK, 2011; pp. 228–237)
- Colding, J. Gren, Å. and Barthel, S. (2020) *The Incremental Demise of Urban Green Space*, *Land*, 9, 162; doi:10.3390/land9050162
- de la Barrera, F., Reyes-Paeckeb, S. and Banzhaf, E. (2016) Indicators for green spaces in contrasting urban settings, *Ecological Indicators*, 62, 212-219.
- Ekkel, E.D. and de Vries, S. (2017). Nearby green space and human health: Evaluating accessibility metrics, *Landscape and Urban Planning*, 157, 214-220.

- Fredman, P. and Hedblom, M. (2015) *Outdoor life 2014: National survey on the Swedish people's outdoor habits*, the Swedish Environmental Protection Agency, Report 6691, Bromma, Sweden. In Swedish.
- Gibbons, P., & Lindenmayer, D. B. (2007). Offsets for land clearing: No net loss or the tail wagging the dog? *Ecological Management & Restoration*, 8(1), 26-31.
doi:<https://doi.org/10.1111/j.1442-8903.2007.00328.x>
- Griffiths, V.F., Bull, J.W., Baker, J., Milner-Gulland, E.-J. (2018) No net loss for people and biodiversity, *Conservation Biology*, 33, 76-87.
- Jim, C.Y. and Chen, W. (2006). Recreation-Amenity Use and Contingent Valuation of Urban Green Spaces in Guangzhou, China. *Landscape and Urban Planning*, 75, 81-96.
- Jim, C.Y. and Chen, Y. (2010). External effects of neighborhood parks and landscape elements on high-rise residential value, *Land Use Policy*, 27, 662-670.
- Kahneman, D., and Tversky, A. (1979). Prospect Theory: An Analysis of Decisions Under Risk, *Econometrica*, 47, 263–291.
- Knetch, J.L., Ryanto, Y.E., Zong, J. (2012). Gain and Loss Domains and the Choice of Welfare Measure of Positive and Negative Changes. *Journal of Benefit-Cost Analysis*, 3, 1-18.
- Latinopoulos, D., Mallios, Z. and Latinopoulos, P. (2016). Valuing the benefits of an urban park project: a contingent valuation study in Thessaloniki, Greece, *Land Use Policy*, 55, 130–141.
- Lo, A.Y. and Jim, C.Y. (2010). Willingness of residents to pay and motives for conservation of urban green spaces in the compact City of Hong Kong. *Urban Forestry and Urban Green*, 9, 113–120.
- Lund Municipality (2015) *Detaljplan för Vipemöllan 38 m fl i Lund, Lunds kommun (Vipeholmsvägen), Granskningshandling PÅ 25/2012a, 1281K-P129*. “Detailed plan for Vipemöllan 38 and more in Lund, Lund Municipality”.
- Lund Municipality (2018) *Fördjupning av översiktsplanen för Öresundsvägen med omnejd – Västerbro, Södra Gunnesbo, Traktorvägen, Pilsåker och Norra Värpinge*. Antagen av kommunfullmäktige 2018-01-25. “Details of superficial plan for Öresundsvägen and surrounding areas, Västerbro, Södra Gunnesbo, Traktorvägen, Pilsåker and Norra Värpinge”.
- McKinnely, L.M. (2002) Urbanization, Biodiversity, and Conservation, *BioScience*, 52, 883-890.
- Miller, A.R. 2001. *Valuing open space: Land economics and neighborhood parks*. Cambridge. MA: Massachusetts Institute of Real Estate.
- Moilanen, A., and Kotiahe. J.S. (2018) Fifteen operationally important decisions in the planning of biodiversity offsets, *Biological Conservation*, 227, 112–120.
- Morancho, A.B. (2003) A hedonic valuation of urban green areas, *Landscape and Urban Planning*, 66, 35–41.

- Panduro, T. and Veie, K. (2013). Classification and valuation of urban green spaces – A hedonic house price valuation, *Landscape and Urban Planning*, 120, 119-128.
- Peroni, G., Andrews, B., Kontoleon, A. and Bateman, I. (2014) The Value of Urban Green Space in Britain: A Methodological Framework for Spatially Referenced Benefit Transfer, *Environmental and Resource Economics*, 57, 251–272
- Plottu, E. and Plottu, E. (2007) The concept of Total Economic Value of environment: A reconsideration within a hierarchical rationality, *Ecological Economics*, 61, 52-61.
- Plummer, M.L. (2009) Assessing benefit transfer for the valuation of ecosystem services, *Frontiers in Ecology and the Environment*, 7, 38-45.
- Sang, Å.O., Knez, I. Gunnarsson, B. and Hedblom, M. (2016) The effects of naturalness, gender, and age on how urban green space is perceived and used, *Urban Forestry and Urban Greening*, 18, 268-276.
- Schipperijn, J., Ekholm, O., Stigsdotter, U.K., Toftager, M., Bentsen, P., Kamper-Jørgensen, F. and Randrup, T.B. (2010). Factors influencing the use of green space: Results from a Danish national representative survey, *Landscape and Urban Planning*, 95, 130-137.
- Statistics Sweden (2020), table entitled "Folkmängden i Sveriges kommuner 1950-2019 enligt indelning 1 januari 2019, excelfile available at www.scb.se/hitta-statistik/statistik-efter-amne/befolkning/befolkningens-sammansattning/befolkningsstatistik/ (Downloaded 2020-04-08) "The population in Swedish municipalities 1950-2019 according to classification 1 January 2019."
- ten Kate, K., Bishop, J. & Bayon, R. (2004) Biodiversity Offsets: Views, Experience, and the Business Case. IUCN, Gland, Switzerland and Cambridge, UK and Insight Investment, London, UK
- Tzoulas, K., Korpela, K., Venn, S., Yli-Pelkonen, V., Kaźmierczak, A., Niemela, J., James, P. (2007) Promoting ecosystem and human health in urban areas using Green Infrastructure: a literature review, *Landscape and Urban Planning*. 81 (3), 167–178.
- Wolch, J.R., Byrne, J. and Newell, J.P. (2014) Urban green space, public health, and environmental justice: The challenge of making cities 'just green enough', *Landscape and Urban Planning*, 125, 234-244.
- Wu, Z., Chen, R., Meadows, M.E., Sengupta, D. and Xu, D. (2019) Changing urban green spaces in Shanghai: trends, drivers and policy implications, *Land Use Policy*, 87, 104080)

About AgriFood Economics Centre

AgriFood Economics Centre provides economic expertise in the fields of food, agriculture, fishing and rural development. The Centre is a cooperation for applied research between the Swedish University of Agricultural Sciences (SLU) and Lund University. The aim is to supply government bodies with a solid scientific foundation supporting strategic and long-term policy choices.

Publications can be ordered free of charge from www.agrifood.se

AgriFood Economics Centre
PO Box 7080
SE-220 07 Lund
SWEDEN

<https://www.agrifood.se>
mail: info@agrifood.se

